



Citation for published version:

Cooper, P 2013, 'Socio-ecological accounting: DPSWR, a modified DPSIR framework, and its application to marine ecosystems', *Ecological Economics*, vol. 94, pp. 106-115. <https://doi.org/10.1016/j.ecolecon.2013.07.010>

DOI:

[10.1016/j.ecolecon.2013.07.010](https://doi.org/10.1016/j.ecolecon.2013.07.010)

Publication date:

2013

Document Version

Peer reviewed version

[Link to publication](#)

NOTICE: this is the author's version of a work that was accepted for publication in *Ecological Economics*. Changes resulting from the publishing process, such as peer review, editing, corrections, structural formatting, and other quality control mechanisms may not be reflected in this document. Changes may have been made to this work since it was submitted for publication. A definitive version was subsequently published in *Ecological Economics*, vol 94, 2013, DOI 10.1016/j.ecolecon.2013.07.010

University of Bath

Alternative formats

If you require this document in an alternative format, please contact:
openaccess@bath.ac.uk

General rights

Copyright and moral rights for the publications made accessible in the public portal are retained by the authors and/or other copyright owners and it is a condition of accessing publications that users recognise and abide by the legal requirements associated with these rights.

Take down policy

If you believe that this document breaches copyright please contact us providing details, and we will remove access to the work immediately and investigate your claim.

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28

**Socio-Ecological Accounting: DPSWR, a Modified DPSIR Framework, and its
Application to Marine Ecosystems**

Philip Cooper^a

^a School of Management, University of Bath, Bath BA2 7AY, UK.

Tel. 44 (0)1225 386544

Fax 44 (0)1225 386473

email: p.cooper@bath.ac.uk

29 **Abstract**

30

31 In contrast to institutionally focussed environmental accounting, socio-ecological accounting
32 frameworks organise information concerned with human-environment interactions at scales
33 relevant to ecosystem change and thus encapsulate information more relevant to ecosystem-
34 based management. The DPSIR (Driver-Pressure-State Impact-Response) framework has
35 been used to identify relevant information in a number of ecosystem contexts but suffers
36 limitations in terms of its definitional clarity and conceptual foundations, which undermine
37 comparability between studies. These limitations are addressed in the DPSWR (Driver-
38 Pressure-State-Welfare-Response) framework, which defines information categories based on
39 a synthesis of concepts in DPSIR and its predecessors so as to more clearly identify the object
40 of measurement in each category and isolate information relating to social systems.
41 Consequently, its categories dealing with social systems are better suited to assessing
42 anthropocentric trade-offs in environmental decision-making, such as through cost-benefit
43 analysis. A conceptual input-output analysis is used to highlight measurement issues
44 connected with the inter-relations between information categories, particularly as regards
45 scale, and the application of the framework is illustrated by reference to issues affecting
46 marine ecosystems included in a Europe-wide study for the European Commission. However,
47 DPSWR's definitions are designed to be sufficiently general as to support application in other
48 ecosystem contexts.

49

50 **Keywords:** socio-ecological accounting; environmental accounting; ecosystem approach;
51 marine ecosystems

52

53
54 **1. Introduction**

55
56 Environmental accounting involves organising and presenting information on interactions
57 between the economy and the environment in a standardised way to support policy
58 development and decision making (UNSD, 2012) and, in the case of corporate entities, as a
59 basis for social accountability (Gray et al., 1995). This practice is generally associated with
60 institutional accounting frameworks, where the scope of accounting is determined by the
61 attribution of environmental resources to socially defined entities. However, indicator-based
62 frameworks which integrate information on associated changes in social (human) and
63 ecological systems with a scope relevant to ecological change constitute the basis for socio-
64 ecological accounting frameworks (Table 1). Thus, they are more appropriate to fulfilling the
65 information needs of integrated management (Grumbine, 1994; McFadden and Barnes, 2009)
66 as embodied in the ecosystem approach (see, for example, CBD, 2000; Environment Canada,
67 2012; EU, 2008).

68

69 **[Table 1 about here]**

70

71 In the simplest conceivable cases, analysis can be confined to a single ecosystem change with
72 a unique anthropogenic cause at a given location. More generally, socio-ecological
73 accounting requires information categorised so as to capture multiple causes of ecosystem
74 change, the nature of that change and the range of effects on social systems in a manner that
75 supports the analysis of the complex interactions among those categories. The DPSIR
76 (Driver-Pressure-State-Impact-Response) accounting framework comprises information
77 categories based on a chain of causal relations that together encompass these phenomena and
78 has been widely adopted in a variety of ecosystem contexts, for example: soil erosion (Gobin

79 *et al.*, 2004); biodiversity loss (Maxim *et al.*, 2009); and, marine and coastal systems (Bowen
80 and Riley, 2003; Cave *et al.*, 2003). However, as explained below, the definitions of this
81 framework's information categories are subject to interpretation and they are not well aligned
82 conceptually with the economic analysis of interactions between social and ecological
83 systems.

84
85 Some flexibility in the interpretation of information categories allows users to adapt the
86 framework to their particular needs but this flexibility needs to be limited to ensure
87 comparability between studies, making the accumulation of knowledge about specific social-
88 ecological interactions more efficient. This issue was highlighted in a recent European
89 Commission-funded project, ELME (European Lifestyles and Marine Ecosystems), which
90 motivated the study reported here. The project involved multidisciplinary teams analyzing a
91 range of ecosystem changes in European seas making it necessary to adopt a standard set of
92 information category definitions that could be uniformly applied in each case, regardless of
93 the type of ecosystem change or sea under investigation, and supporting common indicator
94 measures so that the results could be aggregated. For this purpose, a modified DPSIR
95 framework (DPSWR, Driver-Pressure-State-Welfare-Response) was designed to improve
96 definitional clarity and the alignment of social system categories with the needs of economic
97 analysis. While the DPSWR framework was thus inspired by the needs of the ELME project,
98 and its application is illustrated here in the context of marine ecosystem degradation, the
99 definition of terms is sufficiently general for it to be utilised in other contexts.

100
101 This paper describes the derivation of the DPSWR framework in two stages. First, it
102 critically reviews the DPSIR category definitions to identify definitional and conceptual
103 limitations, supported by comparison with predecessor frameworks which have a common

104 conceptual heritage - DPSIR may be seen as the latest generation in an evolutionary process.
105 In doing so, the paper provides a historical perspective on the origins of current practice,
106 highlighting the risks of conceptual confusion that have arisen. Second, the results of the
107 review are synthesised through modifications to DPSIR that result in the DPSWR framework.
108 A spatial input-output representation of the causal relationships among information categories
109 is employed to discuss measurement issues in the application of the DPSWR definitions. In
110 these sections, eutrophication, primarily from agricultural run-off, is used to illustrate certain
111 points; the subsequent section uses results from a survey of natural scientists in the ELME
112 project to illustrate other applications. Finally, concluding remarks deal with the advantages
113 of the DPSWR framework, how its scope can be reconciled with that of institutional
114 accounting frameworks and aspects of the framework requiring further development.

115

116 **2. Review of DPSIR and Predecessor Frameworks**

117

118 The review is structured around the DPSIR information categories as originally defined by the
119 European Environment Agency (EEA, 1999) or as subsequently modified in its glossary
120 (EEA, 2012), as summarised in Table 2.¹ This table aligns, as far as possible, these categories
121 with those employed in predecessor frameworks extending back to the Stress-Response
122 Environmental Statistical System (S-RESS) proposed by Friend (1979); no earlier
123 frameworks could be identified in the literature.

124

125

[Table 2 about here]

¹ The European Environment Agency is considered an authoritative source as it or its parent body, the European Commission, have been cited as the source for the framework (e.g. Bowen and Riley, 2003) and the framework has been widely applied by the EEA itself (e.g. EEA, 2007) and in projects for the Commission (e.g. Eurostat, 1999).

126 2.1 Overview

127

128 The summary in Table 2 demonstrates the shared heritage of core concepts in terms of the
129 characteristics of social and ecological systems of relevance, but highlights omissions in
130 certain categories, most notably the restriction to environmental effects in the FDES
131 (Framework for the Development of Environment Statistics) and PSR (Pressure-State-
132 Response) frameworks, and a sub-division of its pressure category in the PSR/E (Pressure-
133 State-Response/Effect) framework. Nevertheless, a wide variation in terminology and subtle
134 variations in the precise definition of information categories are apparent. For example, the
135 S-RESS “stressor” and the PSR “indirect pressure” categories both refer to human activities,
136 but they most closely align with the Driver concept reflected in DPSIR which refers to large-
137 scale social “developments”. Thus, there is a common reference to “what people do” but a
138 distinction between specific activities in S-RESS and PSR, and a more summary
139 representation based on trends in DPSIR.

140

141 Certain of the frameworks are not characterised solely by the definitions of the constituent
142 information categories but also by a structuring property: an orientation towards a motivating
143 object of analysis. S-RESS is additionally structured around “stressor activities”, e.g.
144 generation of waste residuals, harvesting activity, while affected systems provide the
145 orientation for FDES, which uses a media, or environmental component, approach (e.g.
146 atmosphere, water) based on an international survey (UN, 1977). Similarly, PSR is orientated
147 around “issues” associated with particular environmental media, e.g. climate change, water
148 resources. However, such structuring properties are more in the nature of guidance for
149 application of the framework rather than integral to the definition of informational categories.
150 The following sub-sections consider each of the DPSIR categories in turn.

151

152 2.2 Driver Category

153

154 Common to the frameworks summarised in Table 2 is the concept that environmental status or
155 change is ultimately the result of, or driven by, humans.² However, there is little consensus
156 around how this force should be defined and measured.

157

158 The EEA definition represents Drivers as “developments” in fundamental, broad-scale aspects
159 of social systems. While this scope is appropriate to some forms of analysis, it does not allow
160 for the concept of some constant aspect of social systems exerting an influence on ecosystems
161 and, moreover, involves information that may be too highly aggregated to elucidate the
162 interactions between social and ecological systems, e.g. in terms of the extent to which they
163 are coupled. To illustrate in the context of eutrophication, in the EAA scheme one of the
164 relevant “developments”, and so a Driver, might be an increasing demand for agricultural
165 output with an associated Pressure being an increased release of eutrophying agents (nitrogen,
166 phosphorus, potassium), i.e. assuming increased use of man-made fertilisers to meet the
167 increased demand. This approach would then fail to capture information on the effect of
168 fertiliser use where there is a constant rate of such use, or indeed varying rates of use where
169 there is no change in demand for agricultural output, and ignores the potential for decoupling
170 in the relationship between fundamental economic forces and pressures. For example,
171 changes in price or technology, or policy measures, may result in decoupling between
172 fertiliser inputs and agricultural output.

² While the EEA glossary refers to “driving force(s)” as opposed to “driver(s)” as used elsewhere (e.g. Cave *et al.*, 2003), there appears to be no conceptual distinction since the EEA uses the terms interchangeably (EEA, 1999; EEA, 2000).

173
174 Applications and reviews of the framework in practice, see for example, Cave *et al.* (2003),
175 Turner *et al.* (2000) and Wieringa (1999), broadly follow the EEA definition and the above
176 interpretation, although Cave *et al.* (2003) and Turner *et al.* (2000) expand on the nature of
177 drivers by specifying types of change or processes (e.g. urbanisation, agricultural
178 intensification). However, the predecessors of DPSIR (Table 2) reinforce the concern that the
179 framework as defined by the EEA focuses the Driver category at too broad a scale. While
180 different terms are employed, a common theme in the definition of “stressor” (S-RESS),
181 “action” (FDES) and “indirect pressures” (PSR and PSR/E) is that they refer to “activities” –
182 the use of fertilisers in the eutrophication example. Thus, they envisage specific actions as
183 giving rise to environmental effects. Furthermore, the PSR/E framework draws a distinction
184 between “underlying” and “indirect” pressures which reconciles the scope of the DPSIR
185 Driver category to those of other frameworks in that “underlying” pressures correspond to
186 broad-scale changes in social system as envisaged in the DPSIR Driver category while
187 “indirect” pressures correspond to specific activities at the source of environmental change as
188 envisaged in these other frameworks: respectively, the demand for agricultural output and the
189 use of man-made fertilisers in the eutrophication case.

190

191 2.3 Pressure Category

192

193 Consistent with its definition of Driver, the definition of Pressure in the EEA scheme refers to
194 “developments”, again suggesting the need for some change in rate to constitute a pressure
195 although, for example, a constant rate of emission can equally well be seen as a pressure on
196 the assimilative capacity of ecosystems.

197

198 This category is problematical in terms of detecting a common theme in the other frameworks
199 that might act as a guiding precedent. S-RESS uses the concept of “stress” but defines this in
200 terms of pressures and FDES does not explicitly include a category corresponding to DPSIR’s
201 Pressure. However, both PSR and PSR/E regard a pressure as being a manifestation of
202 human activities that directly acts on the environment (i.e. as “proximate” or “direct”
203 pressures respectively), the simplest example being emissions. This notion is implicit in the
204 PSIR and DPSIR pressure categories but by their reference to resource and land use also
205 encompasses human activities that are elsewhere treated as the equivalent of drivers, creating
206 a potential conflict between the categories.

207

208 2.4 State Category

209

210 Distinct from the definitions of other categories, EEA’s definition of State refers to
211 measurement (indicators) rather than simply what is the object of measurement but it is
212 unclear whether this has any significance. Furthermore, it is unclear whether its reference to
213 the “condition of *different* environmental compartments and systems” (emphasis added)
214 envisages that a measure of State must summarise various aspects of the environment related
215 to a given issue. Together, these observations indicate the need to define State in terms of the
216 object of measurement (rather than a particular measure) while allowing this object to be that
217 most relevant to the analysis in hand (rather than to “different ... compartments”).

218

219 As regards what is to be measured, the EEA definition of State refers to “condition”,
220 suggesting that relevant information in this category is concerned with status at a given point
221 in time. Similarly, the PSR/E and PSR frameworks envisage a static measure reflecting the
222 stock of the quantity or quality of environmental resources. Such measures may be useful in

223 conveying readily accessible messages about trends over time. However, in terms of relating
224 this category to others in the framework, information on change in State would be needed.
225 Indeed, the PSIR and FDES frameworks are concerned with such change (and in this sense
226 resource flows) in the environment, although FDES allows for other perspectives given its
227 recognition that there may be a direct link between changes in stocks and environmental
228 impacts (UN, 1984, paras. 46 & 47), while the relevant S-RESS category encompasses both
229 “effects” on the natural environment and to stock of available resources.

230

231 2.5 Impact category

232

233 In DPSIR, impacts are expressed in terms of the effects of “environmental quality” on both
234 social and ecological systems. Thus, in Cave *et al.* (2003) and Turner *et al.* (1998) impacts
235 include both ecosystem and human welfare effects, although in the former case they focus on
236 ecosystem effects such as “reduced water quality”.

237

238 The inclusion of effects on ecosystems obscures the boundary between this category and the
239 State category. Referring to the eutrophication example, the envisaged distinction in DPSIR
240 could be between eutrophication as the relevant State and its consequences, such as the effects
241 of hypoxia on particular species, as the Impact. However, this division would introduce an
242 artificial distinction since ecosystem changes attributable to eutrophication would also fall in
243 the State category since they reflect the “condition of different environmental compartments
244 and systems”. Thus, ecosystem changes attributable to eutrophication could be treated as a
245 manifestation of Impact or an aspect of State. Alternatively, if State is taken to represent the
246 availability of eutrophying agents, then Impact would embrace both the extent of

247 eutrophication and its consequences, and thus would be seeking to encapsulate a wide range
248 of information, particularly given its inclusion of consequences for social systems.

249

250 A further issue arises from including effects on social and ecological systems in the same
251 category. This conflates distinct concepts, with their own systems of measurement, and
252 potentially disguises the relationships between them. Moreover, defining the ultimate
253 consequences of anthropogenic ecosystem change in this way complicates comparison with
254 the human activities that gave rise to them and can thus hamper Response decisions.

255 Separating the two forms of Impact to isolate the effects on social systems would make
256 information in this new category comparable with that in the Driver category. With
257 appropriate economic measurement in these categories there is thus direct correspondence
258 with the elements required for cost-benefit analysis. To illustrate with the eutrophication
259 example, say that the use of fertilisers increases agricultural yields and so welfare by A but
260 leads to ecosystem Impacts, measured in biophysical terms as B, which generate external
261 welfare costs of C. Only A and C are directly commensurable and respectively represent the
262 benefits and costs of fertiliser use.

263

264 Support for the notion of isolating impacts on social systems in a separate information
265 category can be drawn from predecessor frameworks that recognise effects on social systems:
266 PSR/E and PSIR provide separate categories for such effects. Similarly, in their applications
267 of DPSIR, both Bowen and Riley (2003) and Atkins et al. (2011) recognise the need for
268 separate recognition of welfare effects.

269

270 2.6 Response Category

271

272 This category is the most uncontroversial in that the frameworks summarised in Table 2 share
273 a common conception of information dealing with human reaction to environmental
274 status/change. Furthermore, none confines the category to management action at a particular
275 level represented by the other categories. Throughout, categories corresponding to Response
276 include, for example, policy measures to reduce Drivers as much as to remediate or mitigate
277 Impacts.

278

279 2.7 Omitted Categories

280

281 Having considered each of the categories included in DPSIR, the question arises as to whether
282 these omit types of information covered in the other frameworks. In this context, the FDES
283 framework includes “natural events” in its “action” category but this is on the basis that
284 human activities can contribute to certain such events (UN, 1984, para. 41). The relation
285 between these activities and events can be catered for through DPSIR’s Driver and State
286 categories; a separate category is unnecessary. Nevertheless, this concept of “natural events”
287 highlights the broader issue that relationships within ecosystems may be moderated or
288 exacerbated by natural conditions or their variability that are exogenous to DPSIR categories,
289 i.e. not the result of human activity. While S-RESS allows recognition of “natural forces” as
290 a source of stress on the environment, alone or in conjunction with anthropogenic stresses
291 (Friend, 1979, p.76), none of the frameworks allocates a category to information on such
292 conditions/variability. Rather, it is implied that these forces are manifested through changes
293 in the relationships between categories, i.e. in the application of the accounting framework.
294 They do not constitute a separate factor in the chain of causal relationships that set the scope
295 of the information categories.

296

297 **3. Modifying DPSIR – the DPSWR Framework**

298

299 The modifications of the DPSIR framework respond to the definitional and conceptual
300 limitations, and incorporate lessons learned from the examination of precedents, as discussed
301 above. In overview, the number of categories is retained to ensure they are the minimum
302 necessary for representing relevant information and their scope is aligned with social or
303 ecological systems. Table 3 summarises the resulting information category definitions
304 employed in the DPSWR framework and the relationships between them are depicted in
305 Figure 1. This figure also highlights their configuration as regards social or ecological
306 systems, reflecting the distinction between phenomena associated with human agency and
307 those associated with conditions or processes in ecosystems. Thus, the framework isolates
308 categories of information relating to social systems.³

309

310 Each of the DPSWR categories is discussed below within a spatially-framed conceptual
311 framework involving serial input-output relationships represented by notional coefficients. It
312 is emphasised that this is a means of making explanations more concrete and conveying
313 conceptual relationships in a more formalised setting so as to highlight issues in
314 measurement; it does not purport to constitute a model. In reality, the relation between two
315 categories may be non-monotonic (e.g. exhibiting threshold effects) and/or dependent on a
316 variable from another category. Nevertheless, this form of description may be used as a basis
317 for identifying elements relevant to specifying models and inform decisions on relevant scales
318 and the selection of variables appropriate to specific applications of the framework.

³ The form of measurement for each information category is not prescribed by the framework but the system division denotes a broad distinction between types of available measurement. Categories associated with ecological systems involve the objective observation of nature; those associated with social systems may be represented by physical measures or socially constructed measures of 'value'. However, a common criterion in the selection of measures is that they be appropriate to the objective of the analysis for which the framework is employed – in accounting theoretic terms, the framework is a means of processing information, which entails a purpose orientation (Chambers, 1966: 162).

319 Furthermore, the specific application will influence for which categories information is
320 required; not all categories will necessarily be pertinent for every analysis.

321

322 3.1 Driver - Pressure

323

324 The key modification in the Driver category is to focus it on human activities that give rise to
325 Pressure on natural systems. Thus, the relation between the two is made more direct,
326 overcoming DPSIR's apparent concentration on large scale and potentially long-term changes
327 in social systems which are at some remove from specific activities that precipitate ecosystem
328 change. However, scope is offered for recognising these broader changes as well as the
329 immediate activities associated with Pressures to facilitate studies concerned with trends in
330 fundamental factor such as population and consumer choice that influence the level and nature
331 of activities giving rise to Pressure. This is achieved by focussing the definition of a Driver in
332 the modified framework on an activity or process within the social system but, where
333 necessary to accommodate broader scale analysis, allows separation between "immediate
334 Drivers" (those proximate to Pressures) and "underlying Drivers" (more closely
335 corresponding to the Driver category in DPSIR). This is similar to the approach of the PSR/E
336 framework in its definition of "pressures" (Table 2).

337

338 **[Table 3 and Figure 1 about here]**

339

340 In applications involving multiple Drivers, organising information on an economic-sectoral
341 basis can help direct attention to which parts of the economy are most salient in respect of the
342 environmental issue at stake. This may be useful in the scoping or prioritisation of analysis in
343 practice, or in highlighting critical areas for policy development. Similarly, where the

344 analysis is concerned with long-term social trends, organising immediate Drivers in this way
345 can highlight where such trends are having most effect. A further benefit of sectoral
346 classification arises where the application is motivated by understanding the range of impacts
347 that given Drivers have. The sectoral scheme is not prescribed as part of the framework as the
348 most appropriate level of detail is dependent on the application. In the ELME project, for
349 example, a high level scheme with nine sectors derived from Eurostat (1999), as shown at the
350 foot of Table 3, was adequate.

351
352 In DPSWR, the Pressure category takes on a broad meaning, representing the mechanism or
353 process that intermediates between human action (Driver) and the relevant ecosystem State.
354 As such, the category has a more abstract definition than in DPSIR and allows various uses
355 dependent on the analytical context. This flexibility aids the applicability of the DPSWR
356 framework to a range of contexts within the constraint of the Driver and State category
357 definitions being more firmly fixed. Indeed, in certain cases the category may be redundant.
358 For example, in the case of an emission from some human activity, it is the resulting increase
359 in concentration of the pollutant in the environment that constitutes the Pressure on the
360 ecosystem measured in the State category. By contrast, if the human activity directly affects
361 the ecosystem of interest, e.g. loss of habitat due to dredging activities, there is no need to
362 specify a Pressure measure. In such cases, the Driver category can still be used to examine
363 alternative actions, e.g. different methods of dredging and disposal of spoil.

364
365 To examine the implications for measurement in DPSWR's Driver-Pressure relationship,
366 consider measures of a range of human activities, a_{xi} , indexed by the suffix x ($x = 1, 2, \dots, X$)
367 at location i ($i = 1, 2, \dots, I$), the set of which is represented by the column vector \mathbf{A}_i of order X
368 $\times 1$. Given the set of activities, \mathbf{A}_i , the resulting exploitation of environmental resources, e.g.

369 emissions, e_{yi} , can be connected through notional coefficients, α_{yxi} , where y indexes the type
 370 of exploitation. Thus, for location i , the set of exploitations may be represented as:

$$371 \quad \alpha_i \mathbf{A}_i = \mathbf{E}_i \quad \dots\dots\dots (1)$$

373
 374 where α_i is a $Y \times X$ matrix the elements of which are the respective coefficients and \mathbf{E}_i is a
 375 column vector of order $Y \times 1$, the elements of which are the measures of total environmental
 376 exploitation of type y from activities in location i , e_{yi} . For example, if $y = 1$ represents the
 377 emission of nitrogen species to water and $x = 1$ represents the amount of fertiliser used for a
 378 given period, α_{11i} represents the rate at which this pollutant is produced by this activity at
 379 location i . Furthermore, the first element in \mathbf{E}_i is the total emission of nitrogen species to
 380 water from activities at location i in that period, which would also include, for example,
 381 atmospheric deposition of nitrogen from activities involving combustion at this location.
 382 Thus, equation (1) encapsulates the essential features of the Driver-Pressure relationship and
 383 highlights the significance of the coefficients, α_{yxi} , which can be used to represent various
 384 aspects of the relationship. Variation in a given coefficient across locations, i , may be used to
 385 represent differences in natural transport processes, i.e. the extent to which a given Driver
 386 activity results in Pressure over a given period, while variation over time at a given location
 387 could be used to represent information on decoupling between Driver activities and Pressures,
 388 for example due to management practices, or a closer coupling due to exogenous changes.

389
 390 The scale selected to define location and the measures of activity depend on the application.
 391 For example, a national government may restrict its attention to activities within its territory
 392 or jurisdiction while a regional environmental authority, transcending national boundaries, is

393 more likely to be concerned with activities at a geographical scale appropriate to a specific
 394 ecosystem change and thus, potentially, with activities in various countries.

395

396 3.2 Pressure - State

397

398 In common with other frameworks, State is defined in general terms by reference to
 399 “attributes” (see Table 3) to allow flexibility in application but is most directly comparable to
 400 the S-RESS and FDES frameworks scope in explicitly providing for both static and dynamic
 401 measures of environmental conditions. To illustrate by reference to the input-output analysis,
 402 consider static measures of the relevant conditions at location j , c_{zj} , where z ($z = 1, \dots, Z$)
 403 indexes the type of environmental condition. As noted above, such static measures may be
 404 compared at different points in time to identify trends but by themselves they do not generally
 405 support analysis of the relationships between State and its adjacent categories, which involve
 406 measurement of change over periods of time. Therefore, where the application involves
 407 relating information across categories, measures of change in conditions, d_{zj} , across periods
 408 comparable with those used for measurement in other categories, are required. Where the
 409 elements of \mathbf{D}_{ij} (a $Z \times 1$ vector) are the measures of change in different types of condition d_{zj} ,
 410 and β_{ij} is a $Z \times Y$ matrix, the elements of which are coefficients linking respective Pressures
 411 arising from activities at i to change in environmental conditions at j , the relationship between
 412 the Pressure and State categories can be represented by:

$$413 \quad \beta_{ij} \mathbf{E}_i = \mathbf{D}_{ij} \quad \dots\dots\dots (2)$$

415

416 For example, if $z = 1$ is taken to represent eutrophication then d_{1j} as the first element in \mathbf{D}_{ij} is a
 417 measure of the change in eutrophication at location j attributable to Pressures from human
 418 activities at location i .

419
 420 While the coefficients β_{ij} in equation (2) represent purely natural processes, they also
 421 represent information on human-ecosystem interactions in that they reflect to what extent a
 422 unit of Pressure from Driver activities at a given location, i , is translated into State change at j .
 423 Comparing coefficients across different locations j indicates the physical distribution of
 424 environmental consequences from those activities. This is an anthropocentric orientation and
 425 applications motivated by State change at a specified location \mathbf{D}_j resulting from activities at
 426 multiple locations can be represented by (notional) summation:

$$\mathbf{D}_j = \sum_i \mathbf{D}_{ij} \dots\dots\dots (3)$$

429
 430 Indeed, in practice, it is the values in \mathbf{D}_j that are most likely to be observed initially, and
 431 analysis is required to identify which Driver locations, i , are most material in terms of relative
 432 values of e_{yi} and β_{ij} coefficients across locations.

434 3.3 State - Impact/Welfare

435
 436 Modifications of DPSIR's State and Impact categories are interrelated. The underlying motive
 437 is to isolate the effects of ecosystem change in terms of social systems from the ecosystem
 438 changes resulting from Pressures. Thus, the Impact category is redefined in DPSWR to cover
 439 only information relating to social system effects (see Table 3) and the ecosystem changes
 440 which DPSIR would have treated as Impacts are dealt with in the State category so that this

441 category encompasses all ecosystem changes other than those which constitute Pressures;
 442 where the boundary lies between them depends on the application context, as discussed in
 443 section 4 below. The renaming of the Impact category as Welfare signifies this change in
 444 scope and improves communication with natural scientists who can find it challenging to
 445 employ the word “impact” as exclusively relating to social systems and instinctively use it to
 446 also encompass ecosystem change.⁴ The DPSWR nomenclature supports this usage in that
 447 “impact” can refer to State and/or Welfare changes.

448
 449 By designating a separate category for information relevant to assessing Welfare effects of
 450 ecosystem change, the DPSWR framework draws a clear distinction that makes the
 451 relationship between them more explicit and highlights the need for human agency in linking
 452 State to Response, since Response is contingent on human perception of values associated
 453 with State. Moreover, this distinction supports comparison of the human causes and effects of
 454 environmental change: the Driver category identifies activities the economic benefit of which
 455 can be assessed next to the costs in the Welfare category. Furthermore, the distribution of
 456 benefits and costs can be revealed in the measurement scheme. Continuing the input-output
 457 representation above, the change environmental conditions at location j can be notionally
 458 translated into welfare effects at location k , w_k , through the $K \times Z$ coefficient matrix, γ_{jk} to
 459 yield a vector of welfare effects, \mathbf{W}_k :

$$\gamma_{jk} \mathbf{D}_j = \mathbf{W}_k \quad \dots\dots\dots (4)$$

462 The locations, k , thus represent where the costs of environmental degradation are borne, while
 463 the locations, i (equation 1) represent those where the activities giving rise to that change took
 464 place and thus where benefits arise. Any mismatch between the two reflects different

⁴I am grateful to Laurence Mee for this insight.

465 distributions of benefits and costs and where these cross institutional, e.g. national, boundaries
466 this raises the question of equity.

467

468 Welfare effects encompass a range of human values as regards environmental change,
469 including non-use value motivated by bequest, existence or ‘anthropocentric intrinsic’ values,
470 that are potentially capable of representation with a common (monetary) metric (Turner,
471 1999). Such monetary representation enables lost benefits, or costs, to be aggregated across
472 types of environmental change and their location, and to be compared with monetary
473 representation of the benefits associated with the Driver activities from which the Welfare
474 effects derive, i.e. cost-benefit analysis. Similarly, such analysis might be applied in
475 assessing the cost of Response actions against the benefit of environmental improvement.
476 However, in common with the other DPSWR categories, the Welfare category is defined
477 conceptually rather than in terms of its measurement base. Consequently, in applications
478 where there is a lack of reliable valuation data, indicator-based measures of relevant criteria,
479 e.g. an increase in morbidity or reduction in employment attributable to State change, may be
480 used to represent the Welfare category. With appropriate weightings, such indicators can be
481 combined to represent multiple criteria in a single measure (i.e. multi-criteria analysis, OECD,
482 1989: 19) that can substitute for an aggregate monetary value for those criteria.

483

484 Through its use of the welfare concept, the DPSWR framework gives a primary role to
485 assessing changes in ecosystem services in terms of the effects on humans, thus aligning it
486 with the information needs of various decision making contexts (Fisher et al., 2009) and
487 policies such as the Marine Strategy Framework Directive (EU, 2008:29). However, the
488 content of the Welfare category can be adapted to other decision-frames such that ‘welfare’ is
489 given a broader meaning as ‘what matters’ through human instrumentality, i.e. human agency

490 in defining decision criteria. For example, where cost-benefit analysis is rejected on
491 economic-theoretic grounds (Gowdy, 2004), under conditions of uncertainty (Perrings, 1991),
492 or, more generally, as an inappropriate basis for environmental decision-making (O'Neill,
493 1997; Sagoff, 2004), the category could be used to encapsulate information on minimum
494 acceptable levels of ecosystem provision in accordance with the precautionary principle or an
495 ethically motivated desire to maintain such a level 'for its own sake' (i.e. 'anthropocentric
496 intrinsic value' in the classification scheme of Turner, 1999:35).

497

498 **4. Illustration in Marine Ecosystems**

499

500 To test the feasibility of the DPSWR framework, natural science teams working in the ELME
501 project were requested to apply the framework in identifying key Drivers and Pressures for
502 the environmental issues with which they were concerned, using the definitions and a slightly
503 extended version of the notes shown in Table 3. Examples derived from their responses are
504 presented in Tables 4 and 5 to illustrate the application of the framework in the marine
505 context. These examples are selected as they involve differing treatments of the same class of
506 item: seagrass loss in the Mediterranean (Table 4) includes various forms of pollution as
507 Pressures while that of chemical pollution in the Northeast Atlantic (Table 5) takes pollution,
508 rather than its consequences for ecosystems, as the State of interest. This reflects the scope of
509 work of the respective teams and illustrates that the treatment of an item as a Pressure or a
510 State variable depends on the application context. A similar flexibility is apparent in policy.
511 The EU's Marine Strategy Framework Directive (EU, 2008) prescribes descriptors of "good
512 environmental status" (in Annex I). Some of these directly correspond to State (e.g. no.1
513 "biological diversity is maintained"), while others may be seen more generally as Pressures

514 (e.g. no.11 “introduction of energy ... is at levels that do not adversely affect the marine
515 environment”).

516

517 **[Tables 4 and 5 about here]**

518

519 Despite the difference in the role of pollution in these cases, both are able to align biophysical
520 phenomena relevant to their scope with Driver and Pressure categories. However, this involves
521 different approaches to the classification of Pressures. In the seagrass case there is a range of
522 types of Pressure (physical, chemical and biological), while for the chemical pollution case
523 this category is used to organise Drivers according to the route of transmission of the
524 pollutants to the marine environment, reflecting the significance of different transport
525 processes as contributors to State change.

526

527 The identification of Drivers and their sectoral classification by the respondents were
528 consistent with the concepts used in the design of the framework despite the relatively simple
529 definition of Driver. Indeed, there is evidence of a flexible interpretation in the responses in
530 that static physical structures such as shipping infrastructure were seen as Drivers although
531 the “activity or process” with which they are associated is their presence, i.e. “being there”.

532

533 **5. Concluding Remarks**

534

535 The DPSWR framework involves a number of modifications to DPSIR which seek to
536 improve the clarity of information category definitions and establish a conceptual foundation
537 for each category that supports its linkage with other categories. By defining Drivers in terms
538 of human activities, a direct link to Pressures is enabled; by expanding the State category to

539 encompass changes in conditions establishes a link between Pressures and changes in human
540 Welfare over a period of time; and by separating such Welfare changes from State changes
541 the boundary between social and ecological systems and the interaction between them are
542 clearly marked. As a result, the DPSWR information categories relating to social systems
543 highlight the link between the human activities that give rise to environmental change,
544 whether as a result of actions in the Driver or Response categories, and the effect of such
545 change on humans. Thus, the framework isolates information relevant to the requirements of
546 cost-benefit analysis and other decision frames insofar as these reflect human, rather than
547 intrinsic, values.

548

549 Modifying DPSIR in the ways embodied in the DPSWR framework was found to be feasible
550 in the project that motivated this study and imports advantages to applications in terms of
551 supporting the identification of representative variables for each category and comparability
552 across studies. Furthermore, the conceptual input-output analysis employed here to represent
553 relationships between categories indicates types of information needed to fully account for
554 interactions between human and ecological systems.

555

556 In common with previous socio-ecological accounting frameworks, DPSWR is less
557 prescriptive than institutional accounting frameworks in its definition of information to be
558 provided but is more comprehensive in its scope. This allows flexibility in its application but
559 results in a richer, more integrated portrayal of human-ecosystem interactions, as is apparent
560 in reconciling this framework with those employed in institutional accounting (Table 1). In
561 the case of macroeconomic accounting, the primary link with DPSWR is through limiting the
562 State category to a scale in accordance with national boundaries, for example UNSD (2012)
563 defines the scope of “environmental assets” by reference to “the economic territory over

564 which a country has control ... including waters and sea-beds within a country's Exclusive
565 Economic Zone" (p.124). Thus, scope is restricted in accordance with anthropocentric criteria
566 which may not correspond to the scales required to fully capture information on ecological
567 change as envisaged in DPSWR. In corporate level accounting, the current practice (as
568 illustrated by the examples in Table 1) is to limit accountability in reporting and management
569 information systems to the immediate results of the entity's Driver activities, its emissions or
570 resource use, corresponding to Pressures in DPSWR. However, where the entity operates at
571 multiple locations, information on Pressures aggregated at the corporate level may be of
572 limited usefulness.

573

574 While socio-ecological accounting frameworks such as DPSWR offer the potential for more
575 detailed policy-relevant information than institutional frameworks, they have thus far been
576 concerned primarily with the definition of relevant types of information. Their further
577 development requires consideration of how to incorporate information on the temporal lags
578 between measures of different categories (e.g. *when* Drivers are manifested in Welfare
579 changes) and the degree of uncertainty in the relationship between information categories, as
580 well as how these relationships may be affected by other changes in future (the importance of
581 which is noted by Heal, 2007, in the context of national accounting). These factors could be
582 material to decisions based on applications of the DPSWR framework.

583

584 **Acknowledgements**

585

586 Funding support from the European Commission under its Framework 6 and 7 programmes
587 for the projects ELME (European Lifestyles and Marine Ecosystems, contract 505576) and
588 KnowSeas (Knowledge-based Management of European Seas, contract 226675) respectively

589 is gratefully acknowledged, as are contributions from scientists participating in these projects,
590 especially Laurence Mee, Emma Jackson and Eva Garnacho. The paper has benefitted from
591 the comments of three anonymous reviewers.

592

593 **References**

594 Atkins, J.P., Burdon, D., Elliott, M., Gregory, A.J., 2011. Management of the marine
595 environment: Integrating ecosystem services and societal benefits with the DPSIR framework
596 in a systems approach. *Marine Pollution Bulletin* 62, 215-226.

597 BFF, 2002. *City Limits: A Resource Flow and Ecological Footprint Analysis of Greater*
598 *London*, Best Foot Forward Ltd., Oxford.

599 Bowen, R.E., Riley, C., 2003. Socio-economic indicators and integrated coastal management.
600 *Ocean & Coastal Management* 46, 299-312.

601 Cave, R.R., Ledoux, L., Turner, K., Jickells, T., Andrews, J.E., Davies, H., 2003. The
602 Humber catchment and its coastal area: from UK to European perspectives. *The Science of*
603 *the Total Environment* 314, 31-52.

604 CBD, 2000. *Ecosystem Approach*, Decision V6 of the Fifth Meeting of the Conference of the
605 Parties to the Convention on Biological Diversity, CBD Handbook (3rd edition), Secretariat of
606 the Convention on Biological Diversity, Montreal.

607 Chambers, R.J., 1966. *Accounting, Evaluation and Economic Behavior*, Prentice-Hall, New
608 Jersey.

609 EEA, 2012. *Environmental Terminology and Discovery Service*, available at

610 <http://glossary.eea.europa.eu/> (last accessed 30 August 2012)

611 EEA, 2007. *Europe's Environment: The fourth assessment*, European Environment Agency,
612 Copenhagen.

613 EEA, 2000. Are we moving in the right direction? Indicators on transport and environment
614 integration in the EU, Environmental issues series No.12, European Environment Agency,
615 Copenhagen.

616 EEA, 1999. Environmental indicators: Typology and overview, Technical report No.25,
617 European Environment Agency, Copenhagen.

618 Environment Canada, 2012. Vision and Guiding Principles, available at
619 <http://www.ec.gc.ca/ap-pa/default.asp?lang=En&n=E41EDD3B-1> (last accessed 30 August
620 2012).

621 EU, 2008. Directive 2008/56/EC of the European Parliament and of the Council (Marine
622 Strategy Framework Directive), Official Journal of the European Parliament, L164/19-40.

623 Eurostat, 1999. Towards Environmental Pressure Indicators for the EU, 1st edition, Eurostat,
624 Luxembourg.

625 Fisher, B., Turner, R.K., Morling, P., 2009. Defining and classifying ecosystem services for
626 decision making. *Ecological Economics* 68, 643-653.

627 Friend, A., 1979. Frameworks for Environmental Statistics: Recent Experience of Statistics
628 Canada, in Rapport, D. and Friend, A. Towards a comprehensive framework for environment
629 statistics: A stress-response approach, Statistics Canada, Ottawa.

630 Gobin, A., Jones, R., Kirkby, M., Campling, P., Govers, G., Kosmas, C., Gentile, A.R., 2004.
631 Indicators for pan-European assessment and monitoring of soil erosion by water.
632 *Environmental Science and Policy* 7(1), 25-38.

633 Gowdy, J.M., 2004. The Revolution in Welfare Economics and Its Implications for
634 Environmental Valuation and Policy, *Land Economics* 80(2), 239-257.

635 Gray, R., Kouhy, R., Lavers, S., 1995. Corporate social and environmental reporting: A
636 review of the literature and a longitudinal study of UK disclosure. *Accounting, Auditing and
637 Accountability Journal* 8(2), 47-77.

- 638 GRI, 2011. Sustainability Reporting Guidelines, version 3.1, Global Reporting Initiative,
639 Amsterdam.
- 640 Grumbine, R.E., 1994. What is ecosystem management? *Conservation Biology* 8, 27–38.
- 641 Hamilton, K., Clemens, M., 1999. Genuine saving in developing countries. *World Bank*
642 *Economic Review* 13(2), 333-356.
- 643 Heal, G., 2007. Environmental accounting for ecosystems. *Ecological Economics* 63, 693-
644 694.
- 645 Lawn, P., 2007. A Stock-Take of Green National Accounting Initiatives. *Social Indicators*
646 *Research* 80, 427-460.
- 647 Maxim, L., Spangenberg, J.H., O'Connor, M., 2009. An analysis of risks for biodiversity
648 under the DPSIR framework. *Ecological Economics* 69, 12-23.
- 649 McFadden, K.W., Barnes, C., 2009. The implementation of an ecosystem approach to
650 management within a federal government agency. *Marine Policy* 33, 156-163.
- 651 OECD, 2003. *OECD Environmental Indicators: Development, Measurement and Use*,
652 Reference Paper, Organisation for Economic Co-operation and Development, Paris.
- 653 OECD, 1993. *OECD Core Set of Indicators for Environmental Performance Reviews*,
654 *Environment Monographs No. 83*, Organisation for Economic Co-operation and
655 Development, Paris.
- 656 OECD, 1989. *Environmental Policy Benefits: Monetary Valuation*, Organisation for
657 Economic Co-operation and Development, Paris.
- 658 O'Neill, J., 1997. Value Pluralism, Incommensurability and Institutions, in Foster, J. (Ed.)
659 *Valuing Nature*, Routledge, London.
- 660 Pearce, D.W., Atkinson, G., 1993. Capital theory and the measurement of sustainable
661 development: an indicator of weak sustainability. *Ecological Economics* 8, 103-108.

- 662 Perrings, C., 1991. Reserved rationality and the precautionary principle: technological change,
663 time and uncertainty in environmental decision making, in Costanza, R. (Ed.) *Ecological*
664 *Economics: The Science and Management of Sustainability*, Columbia University Press, New
665 York.
- 666 Sagoff, M., 2004. *Price, Principle and the Environment*, Cambridge University Press,
667 Cambridge.
- 668 Schaltegger, S., Burritt, R., 2000. *Contemporary Environmental Accounting: Issues, Concepts*
669 *and Practice*, Greenleaf Publishing, Sheffield.
- 670 Schulze, I., Colby, M., 1994. *A Conceptual Framework to Support the Development and Use*
671 *of Environmental Information*, EPA 230 R 94 012, Environmental Statistics and Information
672 Division, Environmental Protection Agency, Washington D.C.
- 673 Turner, R.K., 2000. Integrating natural and socio-economic science in coastal management.
674 *Journal of Marine Systems* 25, 447-460.
- 675 Turner, R.K., 1999. The Place of Economic Values in Environmental Valuation, in Bateman,
676 I.J., Willis, K.G. (Eds.), *Valuing Environmental Preferences*, Oxford University Press,
677 Oxford.
- 678 Turner, R.K., Georgiou, S., Brouwer, R., Bateman, I.J., Langford, I., 2003. Towards an
679 integrated environmental assessment for wetland and catchment management. *The*
680 *Geographical Journal* 169(2), 99-116.
- 681 Turner, R.K., Lorenzoni, I., Beaumont, N., Bateman, I.J., Langford, I.H., McDonald, A.L.,
682 1998. Coastal Management for Sustainable Development: Analysing Environmental and
683 Socio-Economic Changes on the UK Coast. *The Geographical Journal* 164(3), 269-281.
- 684 UN, 1991. *Concepts and Methods of Environment Statistics: Statistics of the Natural*
685 *Environment*, Studies in Methods Series F, No. 57, United Nations Department of
686 International Economic and Social Affairs, Statistical Office, New York.

687 UN, 1984. A Framework for the Development of Environment Statistics, Statistical Papers
688 Series M, No. 78, United Nations Department of International Economic and Social Affairs,
689 Statistical Office, New York.

690 UN, 1977. Survey of Environment Statistics: Frameworks, Approaches and Statistical
691 Publications, Statistical Papers Series M, No. 73, United Nations Department of International
692 Economic and Social Affairs, Statistical Office, New York.

693 UNSD, 2012. System of Environmental-Economic Accounting, Central Framework, United
694 Nations Statistical Division, New York.

695 UNSD, 2003 Integrated Environmental and Economic Accounting, Handbook of National
696 Statistics, Studies in Methods Series F, No. 61 Rev.1, United Nations Statistical Division,
697 New York.

698 Wackernagel, M., Onisto, L., Bello, P., Callejas Linares, A., López Falfán, I.S., Méndez
699 García, J., Suárez Guerrero, A.I., Suárez Guerrero, Ma. G., 1999. National Natural Capital
700 Accounting with the Ecological Footprint Concept. *Ecological Economics* 29, 375-390.

701 Wieringa, K., 1999. Towards integrated environmental assessment support for the European
702 Community's environmental action programme process. *International Journal of Environment
703 and Pollution* 11(4), 525-541.

704

705 Table 1: A typology of environmental accounting frameworks

Perspective	Economic scope/ Accounting entity	Environmental scope	Measurement	Examples
Institutional	Macroeconomic/ Geo-political entity (typically a country)	Attributable (national) resource stocks and flows	Monetary, e.g. environmentally adjusted national income, “environmental asset” values ^a	Satellite national accounts (UNSD, 2003: 450, 2012) Sustainability indicators (Hamilton and Clemens, 1999; Pearce and Atkinson, 1993)
			Physical units, e.g. physical flows between economy and environment	Satellite national accounts (UNSD, 2012)
			Total resource use	Geographical area required to meet entity’s needs
	Microeconomic/ Corporate entities (e.g. firms)	Attributable natural resource use (as source or sink)	Physical units ^c	Sustainability reporting guidelines (GRI, 2011) Internal ecological accounting (Schaltegger and Burritt, 2000: Ch.11)
Socio- ecological	“Mesoeconomic”/ Scope required to account for social system interactions relevant to the environmental scope	As specified for the purpose of the analysis	Indicators representing status or change by information category	DPSIR (EEA, 2007)

706 ^a For an overview, see Lawn, 2007.707 ^b Ecological footprinting has been applied to sub-national entities, e.g. Greater London (BFF, 2002).708 ^c The use of monetary measures in corporate environmental information systems and reporting (e.g. applying
709 external damage estimates to emissions or costing replacement/amelioration) is conceivable but has not been
710 widely taken up in practice.

711 Table 2: Summary of socio-ecological accounting framework information categories and their content
712

Framework	Information Categories				
	Driver	Pressure	State	Impact	Response
DPSIR (EEA, 2012)* (EEA, 1999)**	Driving force** <i>social, demographic and economic developments in societies and the corresponding changes in life styles, overall levels of consumption and production patterns</i>	Pressure** <i>developments in (the) release of substances (emissions), physical and biological agents, the use of resources and the use of land</i>	State (indicator)* <i>indicator of condition of different environmental compartments and systems in physical, chemical or biological variables</i>	(Environmental) impact* <i>impacts on human beings, ecosystems and man-made capital resulting from changes in environmental quality</i>	Response** <i>responses by groups (and individuals) in society, as well as government attempts to prevent, compensate, ameliorate or adapt to changes in the state of the environment</i>
PSIR (Turner 2000; Turner et al., 1998)	Socio-economic drivers <i>urbanisation and transport/trade, agricultural intensification/land-use change, tourism and recreation demand etc.</i>	Environmental pressures <i>land conversions and reclamation, dredging, aggregates and oil and gas extraction, waste disposal etc.</i>	Environmental “state” changes <i>changes in ... fluxes across coastal zones, loss of habitats and biodiversity etc.</i>	Impacts <i>consequential impacts on human welfare via productivity, health, amenity and existence value changes</i>	Policy response options -
PSR/E (Schulze and Colby, 1994)	Pressures		Direct pressures <i>biophysical inputs and outputs that may exert immediate stress on ecosystems</i>	State <i>Ambient conditions and trends Valued environmental attributes</i>	Societal response <i>purposeful actions to address ... ecological, human health or welfare changes or impacts that are considered undesirable</i>
	Underlying <i>social and technological forces that ... drive economic activity</i>	Indirect ^a <i>human activities related to ... improvement of human welfare</i>			
Effects <i>Relationships between two or more variables within any of the pressure, state and response categories</i>					
PSR (OECD, 2003; 1993)	Pressure		State		Response
	Indirect <i>human activities which lead to proximate pressures</i>	Proximate <i>pressures directly exerted on the environment e.g. emissions</i>	<i>quality of the environment and the quality and quantity of natural resources</i>		<i>actions to mitigate, adapt to or prevent human-induced negative impacts on the environment</i>
FDES (UN, 1991; 1984)	Action <i>social and economic activities, natural events</i>		Impact <i>environmental impacts of activities/events</i>		Reaction <i>responses to environmental impacts</i>
S-RESS	Stressor	Stress	Environmental response		Collective and individual

(Friend, 1979)	<i>activities with the potential to degrade the quality of the natural environment, to effect (sic) the health of man, to threaten the survival of species, to place pressure on non-renewable resources, and to deteriorate the quality of human settlement</i>	<i>elements that place pressures on, and contribute to the breakdown of, the natural and man-made environment</i>	<i>observed effects of stress upon natural and man-made environments</i> Stock <i>available resource (cf. flows in other measures)</i>		responses <i>man's reaction to environmental changes</i>
----------------	--	---	--	--	---

713
714
715

^a Also includes “natural processes and factors” which may act alone or together with human actions to create biophysical pressures.

716 Table 3: Definition of DPSWR information categories
717

Information Category	Definition	Commentary
Driver	An activity or process intended to enhance human welfare.	<ul style="list-style-type: none"> ▪ Organising activities into economic sectors assists in directing attention to the most salient areas of the economy.^a ▪ Where necessary the category can be split between: <ul style="list-style-type: none"> - Immediate Drivers: activities proximal to at least one Pressure. - Underlying Driver: population, economic, social and technological factors that influence the level/nature of Immediate Drivers.
Pressure	A means by which at least one Driver causes or contributes to a change in State.	Thus a pressure is a link between a Driver and a change in environmental State, effectively therefore the agent of change. Generally, it is a by-product or an unintended consequence of the Driver activity/process.
State (change)	An attribute or set of attributes of the natural environment that reflect its integrity as regards a specified issue (or change therein).	<p>This definition allows flexibility so that the information or measure used can be tailored to the precise circumstances that are relevant. However, often the most useful information will:</p> <ul style="list-style-type: none"> ▪ relate to the extent to which a system has been subject to disturbance, particularly in terms of ecosystem functionality, and ▪ reflect changes in State over time. <p>Natural (i.e. non-anthropogenic) variability may influence the effect of Pressures on State or change in State.</p>
Welfare	A change in human welfare attributable to a change in State.	<ul style="list-style-type: none"> ▪ “Change” allows for enhancement but generally we are concerned with diminution in welfare. ▪ Welfare is not only affected by changes in use values; it can be affected by changes in nonuse values that people hold (e.g. in respect of general ecosystem functionality or the viability of particular species).
Response	An initiative intended to reduce at least one Impact (State or Welfare change).	In this sense “initiative” is an action that would not have been taken in the absence of an effect on Welfare. It may operate through influencing any of the above but with the intention to ultimately reduce such an effect.

718 ^a For example, the following sectors were used in the ELME project (based on Eurostat, 1999): Agriculture,
719 Energy, Fisheries & Aquaculture (including extraction of biological resources), Household (individual
720 consumption), Industry, Tourism & Recreation, Transport, Urbanisation & Coastal Development.
721

722 Table 4: Summary of drivers and pressures resulting in loss or degradation of seagrass beds in the Mediterranean
 723

Pressure	Driver	Sector^a
Nutrient (nitrogen & phosphorus) release	Fertiliser use; Intensive livestock management	Agriculture
	Release of sewage after low level treatment; Use of inadequate sewage systems (resulting in leakage and storm water overflows)	Urbanisation
	Aquaculture (discharge of waste food, faecal and dead animal waste)	Fisheries
Mechanical disturbance	Dredging and spoil disposal; Propeller and anchor damage from shipping	Transport
	Laying of submarine pipelines and cables; Land claim	Urbanisation
	Use of mobile gears (e.g. trawling and dredging)	Fisheries
	Aggregate extraction	Resource Extraction
	Anchoring of pleasure craft	Tourism & Recreation
Introduced organisms	Accidental release of organisms from aquaria (e.g. <i>Caulerpa taxifolia</i>)	Tourism & Recreation
	Use of mobile fishing gears (leading to spread of <i>Caulerpa</i>)	Fisheries
Contaminants	Land claim using contaminated landfill; Sewage sources, as above	Urbanisation
	Industrial processes (discharge of untreated wastes)	Industry
	Dredging/spoil disposal; Shipping waste disposal (e.g. oil and fuel)	Transport
Physical oceanographic change (increase in turbidity)	Aggregate extraction	Resource Extraction
	Beach replenishment using terrigenous material	Tourism & Recreation
	Land claim; Coastal defence construction/modification; Discharge of sewage particulates	Urbanisation
	Land management, e.g. deforestation (increased deposition of sediment)	Agriculture
	Aquaculture (discharge of waste food, faecal and dead animal waste)	Fisheries
Direct removal	Harvesting for use in agricultural fertiliser production	Fisheries
	Removal to “improve” aesthetics of tourist beaches.	Tourism & Recreation
Anthropogenic structures (producing shading)	Shipping infrastructure: ports, harbours, jetties	Transport
	Land claim; Protection of urban areas/infrastructure through coastal defences	Urbanisation
Chemical oceanographic change	Water abstraction	Extraction
	Untreated waste discharge (increasing chemical oxygen demand)	Industry
	Untreated sewage discharge (increasing chemical and biological oxygen demand)	Urbanisation
Deposition of physical material	Laying of submarine pipelines and cables; Sewage sludge disposal	Urbanisation
	Construction/maintenance of artificial reefs	Tourism & Recreation

724 ^a “Fisheries” includes biological resource utilisation and aquaculture; “Transport” includes maritime traffic and construction/preservation of navigable routes

725 Table 5: Summary of drivers and pressures resulting in chemical pollution in the Northeast Atlantic

726

Pressure	Driver	Sector
Chemical products discharged into rivers, coastal waters or offshore	Disposal of urban waste; Release of sewage after low level treatment; Use of inadequate sewage systems (resulting in leakage and storm water overflows); Coverage of land with urban space and roads (producing run-off water)	Urbanisation
	Use of domestic chemicals and pharmaceuticals (including hormones) present in wastewater	Household
	Industrial processes	Industry
	Aquaculture (leading to release of chemotherapeutics, hormones)	Fisheries
	Shipping waste disposal; Release of antifouling agents	Transport
Atmospheric emissions deposited in rivers, coastal waters or offshore	Use of pesticides/herbicides (contaminating run-off water)	Agriculture
	Industrial processes	Industry
	Operation of terrestrial vehicle and ship engines	Transport
	Combustion for electricity generation	Energy
Chemicals accidentally released into rivers, coastal waters or offshore	Oil and gas refinery processes; Flaring by oil and gas production platforms	Energy
	Use of pesticides/herbicides	Agriculture
	Industrial processes	Industry
	Transport of oil/chemicals	Transport
	Oil production processes	Energy

727

728

Figure 1: Summary of DPSWR definitions and relationships

